



Contents lists available at ScienceDirect

# International Journal of Hygiene and Environmental Health

journal homepage: [www.elsevier.com/locate/ijheh](http://www.elsevier.com/locate/ijheh)

## Urinary concentrations of polycyclic aromatic hydrocarbons in Israeli adults: Demographic and life-style predictors



Hagai Levine<sup>a,\*</sup>, Tamar Berman<sup>b</sup>, Rebecca Goldsmith<sup>b</sup>, Thomas Göen<sup>c</sup>, Judith Spungen<sup>b</sup>, Lena Novack<sup>d</sup>, Yona Amitai<sup>e</sup>, Tamar Shohat<sup>f</sup>, Itamar Grotto<sup>b,d</sup>

<sup>a</sup> Braun School of Public Health and Community Medicine, Hebrew University-Hadassah and The Hebrew University Center of Excellence in Agriculture and Environmental Health, Jerusalem, Israel

<sup>b</sup> Public Health Services, Ministry of Health, Jerusalem, Israel

<sup>c</sup> Institute and Outpatient Clinic of Occupational, Social and Environmental Medicine, Friedrich-Alexander-Universität Erlangen-Nürnberg, Erlangen, Germany

<sup>d</sup> Faculty of Health Sciences, Ben-Gurion University of the Negev, Beer-Sheva, Israel

<sup>e</sup> Department of Management, Bar Ilan University, Ramat Gan, Israel

<sup>f</sup> Israel Center for Disease Control, Ministry of Health, Ramat Gan, Israel

### ARTICLE INFO

#### Article history:

Received 29 July 2014

Received in revised form

11 September 2014

Accepted 13 September 2014

#### Keywords:

Biomonitoring

Polycyclic aromatic hydrocarbons

Smoking

Ethnicity

Urinary metabolites

Diet

### ABSTRACT

Polycyclic aromatic hydrocarbons (PAHs) are ubiquitous environmental pollutants associated with adverse health outcomes, including cancer, asthma, and reduced fertility. Because data on exposure to these contaminants in Israel and the Middle East are very limited this study was conducted to measure urinary levels of PAHs in the general adult population in Israel and to identify demographic and life-style predictors of exposure.

We measured concentrations of five PAH metabolites: 1-hydroxypyrene (1OH.pyrene) and four different hydroxyphenanthrenes (1-hydroxyphenanthrene, 2-hydroxyphenanthrene, 3-hydroxyphenanthrene, 4-hydroxyphenanthrene), as well as cotinine in urine samples collected from 243 Israeli adults from the general population. We interviewed participants using structured questionnaires to collect detailed demographic, smoking and dietary data. For over 99% of the study participants, urinary concentration of at least one of the PAHs was above both the limit of detection (LOD) and the limit of quantification (LOQ). All PAHs were significantly correlated ( $\rho = 0.67\text{--}0.92$ ). Urinary concentration of hydroxyphenanthrenes, but not 1OH.pyrene, was significantly higher among Arabs and Druze study participants ( $N = 56$ ) compared to Jewish participants ( $N = 183$ ). For 4-hydroxyphenanthrene, concentration in Arabs and Druze was 1.95 (95% CI 1.50–2.52) that of Jews, after controlling for creatinine, age and cotinine levels. Urinary concentrations of all PAHs were significantly higher among current smokers or participants with higher cotinine levels and increased significantly with smoking frequency. While PAHs concentrations were not associated with cotinine concentrations in nonsmokers in the overall study population, PAHs concentration was significantly higher among nonsmoking Jews with cotinine  $\geq$ LOQ ( $1 \mu\text{g/L}$ ), which represents exposure to environmental tobacco smoking, compared to nonsmoking Jews with cotinine concentrations  $<$ LOQ, with the highest ratio for 1OH.pyrene ( $Ratio = 2.38$ , 95% CI 1.47–3.85). Among nonsmoking Arabs and Druze, higher hydroxyphenanthrenes concentrations were found for those consuming grilled food once a month or more. For 3-hydroxyphenanthrene, concentration in those consuming grilled food once a month or more was 2.72 (95% CI 1.01–4.98) times that of those consuming grilled food less than once a month or not at all, after controlling for creatinine, age and cotinine levels.

In conclusion, we found that the general adult population in Israel is widely exposed to PAHs. Exposure differed by ethnic sub-groups both in magnitude and sources of exposure. The finding of higher exposure among Arabs and Druze highlights disparities in environmental exposures across subpopulations and suggests that further research and preventive measure are warranted to reduce PAHs exposure and associated health outcomes, especially in the Arab population in the Middle East.

© 2014 Elsevier GmbH. All rights reserved.

\* Corresponding author. Tel.: +972 2 6777452; fax: +972 2 6431086.

E-mail addresses: [hlevine@hadassah.org.il](mailto:hlevine@hadassah.org.il), [tohagai@bezeqint.net](mailto:tohagai@bezeqint.net) (H. Levine), [tamar.berman@moh.health.gov.il](mailto:tamar.berman@moh.health.gov.il) (T. Berman), [rivka.goldsmith@moh.health.gov.il](mailto:rivka.goldsmith@moh.health.gov.il) (R. Goldsmith), [Thomas.Goen@ipasum.med.uni-erlangen.de](mailto:Thomas.Goen@ipasum.med.uni-erlangen.de) (T. Göen), [jhsungen@gmail.com](mailto:jhsungen@gmail.com) (J. Spungen), [novack@bgu.ac.il](mailto:novack@bgu.ac.il) (L. Novack), [yonaamitai89@gmail.com](mailto:yonaamitai89@gmail.com) (Y. Amitai), [tamar.shohat@icdc.health.gov.il](mailto:tamar.shohat@icdc.health.gov.il) (T. Shohat), [itamar.grotto@moh.health.gov.il](mailto:itamar.grotto@moh.health.gov.il) (I. Grotto).

<http://dx.doi.org/10.1016/j.ijheh.2014.09.004>

1438-4639/© 2014 Elsevier GmbH. All rights reserved.

## Introduction

Polycyclic aromatic hydrocarbons (PAHs) are a group of over 100 chemicals that are generated during incomplete combustion of organic material, such as in coal-fired power plants and residential heating (Boström et al., 2002). PAHs are ubiquitous persistent but non bioaccumulative compounds (Bergman et al., 2013). These substances generally are found in the environment as complex mixtures rather than as single compounds. In the general population, exposure to PAHs can potentially occur via numerous exposure pathways, including active and secondhand smoking (also called environmental tobacco smoke), exposure in indoor and ambient air, and consumption of grilled and smoked foods (CDC, 2013; Phillips, 1999; Suwan-ampai et al., 2009).

Several adverse health outcomes in humans were found to be associated with exposure to PAHs. Evidence indicates that exposure to PAHs may result in reproductive, developmental, hematologic, cardiovascular, neurologic, and immunologic toxicities as well as cancer (Straif et al., 2005). In parallel, new evidence is gathering regarding other harmful effects. For example, a recent study found high prenatal PAHs exposure to be associated with symptoms of anxiety/depression and attention problems (Perera et al., 2012).

Exposure to PAHs is not generally monitored in human biomonitoring surveys, so data on magnitude and predictors of PAHs exposure of the general population are limited, mostly available from the US (CDC, 2013) and Germany (Schulz et al., 2007). 1-Hydroxypyrene (1OH-pyrene) is widely considered as an appropriate biomarker for exposures to PAHs on the basis that pyrene is rapidly distributed, metabolized and eliminated from the body (Aquilina et al., 2010).

Over the last decades, Israel, located in the Middle-East region, has undergone a transition from a developing country into a modern life-style, industrial state with a unique multi-ethnic and multi-cultural society. While ~80% of Israelis are Jews, there is a large minority of Israeli Arabs and Druze who are distinct by cultural, socio-demographic and life-style factors, with higher rates living in rural areas.

The goal of the current study was to measure urinary levels of various PAHs in the general adult population in Israel and to determine the demographic, smoking and dietary predictors of exposure, in order to support evidence based public health policy.

## Materials and methods

### *Study design, settings and participants*

The current study is based on the Israeli Biomonitoring Study which was a cross-sectional study on exposure of Israeli adults from the general population to environmental chemicals and/or their metabolites, measured in urine samples. The primary objective of the biomonitoring study was to provide information on exposure to environmental chemicals in Israel in order to support public health policy. Aims and methods of the biomonitoring study are further detailed in our previous publications (Berman et al., 2013a,b).

The eligible population included Israeli adults, aged 20–74, aiming to represent the Israeli non-institutionalized adult population. Recruitment, interviewing and sampling took place between February and June 2011. The potential sample size was 300 individuals, assuming up to 20% of missing data, incomplete questionnaires and invalid urine samples, to reach a planned sample size of 250 individuals. The parameters for defining the sample were selected so as to represent the population distribution of urban versus rural dwelling (with urban defined as population more than 2000) and the two major ethnic groups in Israel (Jews and Arabs) as well as wide geographical representation. Overall, 20 cities/towns were selected, with 4 within the Arab sector (3 urban, 1 rural) and

16 within the Jewish sector (15 urban, 1 rural), representing the relative proportion of the ethnic groups in Israel. In each city/town, interviewers were requested to interview 15 people. Within each city/town, interviewers were required to select 5 separate areas. Within each area, recruitment was done by “knocking on doors” and interviewing those who met the inclusion criteria and agreed to participate, including providing a urine sample. The inclusion criteria were age (20–74) and ability to answer the questionnaire in Hebrew or Arabic. The response rate was 29%, excluding individuals not eligible for the study and individuals not at home at the time of the visit. People refusing participation were replaced by “knocking on the next door”. Participants were not targeted for specific health status and were not included or excluded on the basis of their potential for low or high exposures to environmental chemicals including tobacco exposure. Of 249 participants eventually included in the biomonitoring study, 243 were examined for PAHs urinary levels.

The study was conducted in accordance with the ethical principles of the Declaration of Helsinki. The study protocol was reviewed and approved by the Sheba Tel Hashomer Helsinki Committee. Written informed consent was obtained for all respondents. Participation in the study was voluntary. At the time of recruitment participants received a note explaining that they would receive individual results on urinary concentrations of environmental contaminants if they requested them during their interview or if their individual urinary metabolite results were unusual (more than 10 times the 90th percentile for the study population). All individuals receiving results were invited to contact the study coordinator for additional information. All analysis of data for the study was conducted without details on the identity of the participants.

### *Data sources and variables*

Study participants were interviewed using a structured questionnaire. The interviews were administered by trained interviewers. The interview consisted of a health and lifestyle questionnaire, including smoking habits, demographic questionnaire, a 24 h dietary recall and a food frequency questionnaire. All completed questionnaires were returned to the Israel Center for Disease Control for data entry, quality assurance and analysis. Socio-demographic and personal variables included age (analyzed as a continuous variable, and also grouped as “younger”, 20–44 years, and “older” 45–74 years), sex (males/females), ethnicity (Jews/Arabs and Druze), urbanicity (urban/rural), country of birth (Israel/other), education (lower education [below high school qualification]/high school education and above), marital status (single/married/divorce/widowed) and average monthly household income in shekels (<5000/5000–10,000/>10,000).

Dietary questions included vegetarian status, grilled food consumption and smoked food consumption. Questions used were: “Do you define yourself as vegetarian or vegan?”, “How frequently do you eat food that was cooked on a grill or over open fire?”, “How frequently do you eat smoked products, e.g. sausage, fish or smoked cheese?”. Based on the first question, vegetarian status was classified as vegetarian, vegan or not-vegetarian. Based on the second and third question, participants were classified by grilled or smoked food consumption (No or less than once per month vs. Yes, for each).

Smoking status was based on self-report, validated by our previous report on cotinine levels (Levine et al., 2013). Questions used for active tobacco smoking status were: “Do you currently smoke, including hookah (water pipe)?”, “What do you currently smoke, or what did you smoke before? (cigarettes, cigars, pipe, waterpipe, other)”, “How many cigarettes do you smoke per day or per week?” Based on the first question, participants were classified as tobacco smokers (of any kind) or nonsmokers. Active smoking variables were smoking type (cigarettes/waterpipe only) and cigarette

smoking frequency (<10 cigarettes/day, 10–20 cigarettes/day, >20 cigarettes/day). Some of the participants categorized as cigarette smokers also smoked cigars or were waterpipe smokers.

Urine spot samples were collected in 120-mL urine specimen containers. All urine samples were maintained at below 4 °C for a maximum of 24 h until they were transported to the Sheba Medical Center at Tel Hashomer. Urine samples were aliquoted at Sheba Medical Center and frozen at –20 °C. Within four months of collection, urine samples were shipped to the University of Erlangen–Nuremberg in Germany on dry ice (–70 °C), where they were analyzed.

Laboratory analyzes of PAHs, cotinine and creatinine were performed at the Institute and Outpatient Clinic of Occupational, Social and Environmental Medicine, University Erlangen–Nuremberg in Germany using the following methods. The determination of hydroxylated PAHs (1-hydroxypyrene (1OH.pyrene), 1-hydroxyphenanthrene (1OH.phen), sum of 2- and 9-hydroxyphenanthrene (2OH.phen), 3-hydroxyphenanthrene (3OH.phen), 4-hydroxyphenanthrene (4OH.phen) was performed using high-performance liquid chromatography with fluorescence detection (HPLC-FD) (Hemat et al., 2012). Clean-up took place using an online extraction procedure with a copper phthalocyanine-modified silica gel phase (pre concentration column). Calibration standards were prepared by spiking pooled non-smoker urine with the hydroxylated PAHs. The limits of quantification (LOQ) were 16 ng/L urine for 1OH.phen, 4 ng/L for 2OH.phen, 5 ng/L for 3OH.phen, 8 ng/L for 4OH.phen and 12 ng/L for 1OH.pyrene (Feldt et al., 2014). Aliquots of quality control material prepared in human urine were analyzed in each series to verify the comparability of the analyses. Analytical accuracy was demonstrated for 1OH.pyrene (as well as cotinine) by the successful participation in the proficiency test of the German External Quality Assessment Scheme (G-EQUAS) (Göen et al., 2012).

Cotinine in urine was determined using a gas chromatography mass spectrometry procedure validated and published by the working group “Analyzes in biological materials” (Müller, 2003). In brief, cotinine was extracted from the urine using dichloromethane and quantified after gas chromatographic separation by mass spectrometry in single ion monitoring mode (Eckert et al., 2011), as further detailed in our previous report on cotinine levels (Levine et al., 2013). LOQ was 1 µg/L for cotinine. Limit of detection (LOD) and LOQ were calculated based on a signal-to-noise ratio of 1 to 3 and 1 to 6, respectively. Urinary analyte concentrations were provided in units of µg/L. These concentrations were divided by urinary creatinine concentrations (g creatinine/L urine) to generate creatinine-adjusted analyte concentrations. Creatinine in urine was determined photometrically as picrate according to the Jaffé method (Larsen, 1972).

#### Statistical methods

We calculated medians, geometric means, 95% confidence intervals and maximum values for urinary analyte concentrations and for creatinine-adjusted analytes for each of the PAHs metabolites. Concentrations below the limit of quantification (LOQ) for an analyte were replaced by the limit of detection (LOD) for that analyte, and concentrations below the LOD were replaced with the LOD divided by the square root of 2 (Hornung and Reed, 1990).

The main outcome variable was geometric mean (GM) of PAHs concentration. We also reported the proportion  $\geq$ LOQ. Associations between continuous variables such as PAHs metabolites and cotinine were assessed using Spearman correlation tests, overall, and stratified by ethnicity.

As a part of a univariate analysis we compared geometric means of urinary PAH metabolites not adjusted to creatinine of demographic, dietary and smoking subgroups using a ratio t-test

procedure on a lognormal distribution. Analysis of smoking subgroups included smoking habits among smokers and cotinine levels by a cut-off  $\geq 4$  µg/L (for overall analysis) and  $\geq 1$  µg/L (for nonsmokers subgroup analysis) representing higher active and/or passive tobacco exposure (Heinrich et al., 2005). Differences between subgroups were also assessed based on ratios obtained from univariate linear regression, with subgroup indicator as an independent factor and log-transformed PAHs as a dependent variable. We further applied an antilog function to a linear effect of factors examined, resulting in a multiplicative effect (ratio). A multivariable analysis was performed in a similar way to analyze differences by ethnicity, controlling for significant predictors for at least one PAH, which were creatinine, cotinine and age as continuous variables. Similar analyses were used for subgroup analyses, stratified by ethnicity and by smoking status. *p* Value <0.05 was considered statistically significant. We used SAS 9.2 for all the analyses.

## Results

### Characteristics of the study population

The 243 study participants (51% males) were composed of 183 (77%) Jews, 56 (23%) Arabs and Druze and 4 with other ethnicity (Table 1). 66% of the study population was aged 20–44 years (Mean age = 38.5, SD = 12.6). Participants in the study came from five geographic regions in Israel: Northern (*N* = 15), Haifa (*N* = 69), Central (*N* = 49), Tel Aviv (*N* = 60), and Southern (*N* = 50).

### Urinary concentrations of PAHs metabolites

Urinary PAHs were detected at levels >LOQ for over 90% of study participants for all metabolites, except 4OH.phen (63%), which was present at the lowest concentrations (Table 2). 1OH.phen and 1OH.pyrene were the PAHs metabolites present at the highest concentrations, followed by 3OH.phen and 2OH.phen. All metabolites (including 1OH.pyrene and each of the hydroxyphenanthrenes) were significantly correlated with each other ( $\rho = 0.67$ – $0.92$ ,  $p < 0.001$ ; strongest for 2OH.phen and 3OH.phen;  $\rho = 0.92$ ) (Table 3). After stratification for ethnicity, correlations of PAHs were remarkably similar for Jews or Arabs and Druze (data not shown).

### Demographic and dietary predictors

Urinary concentrations of 1OH.phen, 2OH.phen and 4OH.phen were significantly higher among Arabs and Druze compared to Jews (Table 1). Levels of all PAHs metabolites tended to be higher for 20–44 years age group compared to 45–74 years, but reached statistical significance only for 1OH.pyrene (GM = 198 ng/L vs. 130 ng/L;  $p = 0.008$ ). Urinary concentrations of all hydroxyphenanthrenes, but not 1OH.pyrene, were significantly higher among those residing in rural compared to urban areas. In the univariate analysis, no significant differences for any of the PAHs were found by gender, body mass index, educational level, country of birth, marital status, household income or dietary habits (grilled food consumption, smoked food consumption, vegetarian status).

In a multivariable model, adjusted for cotinine, creatinine and age (Table 4), all urinary hydroxyphenanthrenes but not 1OH.pyrene were significantly higher among Arabs and Druze compared to Jews, most notably for 1OH.phen (ratio = 1.87; 95% CI 1.44–2.42;  $p < 0.001$ ) and 4OH.phen (ratio = 1.95; 95% CI 1.50–2.52;  $p < 0.001$ ). Of note, associations were accentuated after adjustment, especially for cotinine. Results were materially unchanged (data not shown) after adjustment for other variables such as gender or body mass index.

**Table 1**  
Participants' characteristics and PAHs urinary concentrations geometric mean (ng/L), Israeli adults, 2011 (N=243).

Characteristics <sup>a</sup>	% (n/N)	1OH.pyrene			1OH.phen			2OH.phen <sup>b</sup>			3OH.phen			4OH.phen <sup>c</sup>		
		GM	95% CI	p-Value	GM	95% CI	p-Value	GM	95% CI	p-Value	GM	95% CI	p-Value	GM	95% CI	p-Value
Sex																
Male	51.4 (125/243)	186	152;228	0.269	184	151;224	0.647	98	83;116	0.373	144	117;177	0.164	28	23;35	0.832
Female	48.6 (118/243)	157	127;196		197	159;245		88	74;104		118	96;144		29	24;36	
Age, years																
20–44	65.8 (160/243)	198	167;235	<u>0.008</u>	210	178;249	0.059	100	86;115	0.097	143	119;170	0.095	30	25;36	0.483
45–74	34.2 (83/243)	130	98;172		157	119;206		81	65;100		110	86;141		27	23;34	
BMI																
<25	58.3 (129/221)	186	153;227	0.087	190	157;230	0.618	90	76;107	0.876	133	109;162	0.538	28	23;34	0.828
≥25	41.7 (92/221)	141	108;184		176	136;227		92	76;112		121	94;155		29	23;37	
Ethnicity																
Jews	76.6 (183/239)	171	143;204	0.972	168	142;200	<u>0.004</u>	84	73;97	<u>0.005</u>	124	104;146	0.227	27	25;36	<u>0.004</u>
Arabs and Druze	23.4 (56/239)	170	127;228		277	214;360		127	102;157		153	114;204		43	23;34	
Residence																
Urban	88.1 (214/243)	168	144;198	0.497	178	152;207	<u>0.011</u>	87	76;99	<u>0.003</u>	120	103;140	<u>0.002</u>	27	23;31	<u>0.003</u>
Rural	11.9 (29/243)	197	128;305		316	209;477		152	113;206		241	168;346		52	33;83	
Education level																
Lower education	21.3 (51/239)	151	105;216	0.320	183	124;268	0.742	92	68;125	0.923	125	84;186	0.700	32	23;45	0.506
High school and above	78.7 (188/239)	181	154;214		194	165;227		94	82;107		134	115;156		28	24;33	
Grilled food consumption																
No or <1/month	69.0 (167/242)	177	147;213	0.513	193	162;230	0.731	93	28;314	0.897	134	112;160	0.581	30	25;36	0.441
≥1/month	31.0 (75/242)	159	124;205		183	140;239		92	74;114		123	95;159		27	20;35	
Smoked food consumption																
No or <1/month	74.5 (178/239)	171	143;205	0.933	187	157;224	0.693	91	78;106	0.479	130	109;154	0.781	28	24;34	0.554
≥1/month	25.5 (61/239)	174	131;231		200	155;224		101	82;123		136	103;179		31	23;42	
Vegetarian status																
Vegetarian	2.9 (7/239)	84	21;333	Ref.	105	28;399	Ref.	51	25;104	Ref.	93	42;207	Ref.	17	7;45	Ref.
Vegan	4.6 (11/239)	175	80;382	0.202	226	111;461	0.167	106	63;178	0.112	117	47;291	0.682	45	22;92	0.083
Non-vegetarian	92.5 (221/239)	176	151;206	0.104	195	168;227	0.161	94	82;106	0.098	132	114;154	0.427	29	25;34	0.233

<sup>a</sup> There were no significant differences by country of birth, household income or marital status.

<sup>b</sup> 2OH.phen is the sum of 2- and 9-hydroxyphenanthrene.

<sup>c</sup> N = 241 for 4OH.phen due to 2 missing.

**Table 2**  
PAHs metabolites concentrations in urine (ng/Liter and ng/g creatinine), Israeli adults, 2011(N=243).

Metabolite	LOQ (ng/L)	% >LOQ	Unadjusted (ng/L)					Creatinine adjusted (ng/g)			
			Median	Maximum	Geometric mean	Lower 95% CI	Upper 95% CI	Median	Geometric mean	Lower 95% CI	Upper 95% CI
1OH_pyrene	12	91	204.20	2690.42	171.64	147.94	199.15	144.51	137.07	122.80	153.91
1OH_phen	16	90	218.84	2584.57	190.35	164.66	220.05	149.05	152.02	135.21	170.91
2OH_phen <sup>a</sup>	4	99	94.75	1050.10	92.91	82.42	104.72	69.83	74.19	67.22	81.89
3OH_phen	5	97	140.65	1887.11	130.56	113.10	150.73	92.67	104.27	92.81	117.14
4OH_phen <sup>b</sup>	8	63	31.92	358.56	28.92	25.07	33.37	23.52	23.13	20.20	26.48

<sup>a</sup> 2OH\_phen is the sum of 2- and 9-hydroxyphenanthrene.

<sup>b</sup> N=241 for 4OH\_phen due to 2 missing.

**Table 3**  
Spearman correlation coefficients between PAHs metabolites and cotinine (unadjusted to creatinine)<sup>a</sup>, Israeli adults, 2011(N=243).

	1OH_pyrene	1OH_phen	2OH_phen <sup>b</sup>	3OH_phen	4OH_phen	Cotinine	Cotinine (Jews)	Cotinine (Arabs and Druze)
1OH_pyrene	1.000	0.834	0.786	0.836	0.669	0.386	0.443	0.196
1OH_phen		1.000	0.900	0.896	0.835	0.380	0.481	0.147
2OH_phen <sup>b</sup>			1.000	0.923	0.850	0.399	0.499	0.179
3OH_phen				1.000	0.804	0.484	0.556	0.308
4OH_phen					1.000	0.402	0.496	0.225

<sup>a</sup> *p*-Value for all correlations coefficients were <0.001, except for correlations between PAHs and cotinine among Arabs and Druze (*n*=56) [*p*>0.05, except for 3OH\_phen; *p*=0.02].

<sup>b</sup> 2OH\_phen is the sum of 2- and 9-hydroxyphenanthrene.

When urbanicity was introduced to the multivariable model, it was not significant, while ethnicity remained significant for all hydroxyphenanthrenes. In a subgroup univariate analysis, among nonsmokers Arabs and Druze, hydroxyphenanthrenes levels were significantly higher (ratios of 2.0–2.6) among those residing in rural areas (19.5%) compared to those residing in urban areas, but no such difference by urbanicity was observed among nonsmoking Jews (2.0% residing in rural areas). In a subgroup univariate analysis, consumption of grilled food once a month or more often was a common (75.6% of the population) and significant risk factor for higher 1OH\_phen, 2OH\_phen and 3OH\_phen levels among nonsmoking Arabs and Druze (*N*=41) but not among nonsmoking Jews (64.7% consumed grilled food once a month or more often). This finding was most notable for 3OH\_phen (ratio=2.87; 95% CI 1.37–6.02, *p*=0.008), slightly weakened after adjustment for cotinine, creatinine and age as continuous variables (ratio=2.72; 95% CI 1.01–4.98, *p*=0.05).

#### Smoking (including cotinine levels) predictors

Cotinine was significantly correlated with any of PAHs ( $\rho=0.38$ – $0.48$ ) [Table 3]. However, correlations between cotinine and PAHs were stronger and significant for Jews ( $\rho=0.44$ – $0.56$ ), while weaker and non-significant ( $\rho=0.15$ – $0.23$ , except for 3OH\_phen;  $\rho=0.31$ , *p*=0.02) for Arabs and Druze. Urinary concentrations of all PAHs were significantly higher for current smokers compared to non-smokers as categorized based on self-reporting, as well as by cotinine urinary concentrations (Table 5). Among the group of waterpipe only smokers (*N*=13), 1OH\_pyrene levels were relatively high (GM=205 ng/L vs. GM=135 ng/L among nonsmokers) and not significantly different from cigarette smokers (GM=233 ng/L), while hydroxyphenanthrenes levels among waterpipe smokers were similar to nonsmokers. For all PAHs, concentrations increased monotonically with increasing number of cigarettes smoked per day (GM=429 ng/L for 1OH\_phen among smokers >20 cigarettes/day) with stronger associations observed for hydroxyphenanthrenes than for 1OH\_pyrene.

After stratification for ethnicity (Tables 6a and 6b), differences in PAHs by cotinine levels were stronger for Jews (GM doubled or more for participants with higher cotinine levels) than for

Arabs and Druze (GM 1.2–1.6 times higher for participants with higher cotinine levels). Among Jews, urinary concentrations of all PAHs were significantly higher for current smokers compared to non-smokers (GM 1.7–2.0 times higher). Monotonic increase in PAHs concentrations by number of cigarettes smoked per day was observed for Jews, while for Arabs and Druze, concentration did not increase for the five smokers >20 cigarettes/day (for example, for 1OH\_pyrene, GM=403 ng/L for 10–20 cigarettes/day compared to GM=253 ng/L for >20 cigarettes/day). PAHs levels were significantly higher (except 4OH\_phen) among nonsmoking Jews with cotinine concentrations  $\geq$ LOQ (1  $\mu$ g/L), with highest ratio for 1OH\_pyrene (*R*=2.38, 95% CI 1.47–3.85; *p*=0.001). In contrast, PAHs levels did not differ by cotinine levels among nonsmoking Arabs and Druze.

#### Discussion

This study, the first human biomonitoring study of PAHs exposure among general adult population in the Middle East region, shows that the general population in Israel is widely exposed to polycyclic aromatic hydrocarbons. Individual PAH metabolites were highly correlated among each other, suggesting common sources of exposure and similar metabolic pathways (Li et al., 2008). PAHs were associated with cotinine levels and with smoking habits, including smoking frequency. Levels of hydroxyphenanthrenes, but not 1OH\_pyrene, were significantly higher among the Arabs and Druze study participants compared to the Jewish population, after controlling for creatinine, age and cotinine levels. These populations differed by predictors of PAHs exposure among nonsmokers (all or some metabolites): environmental tobacco smoking (as measured by cotinine levels) for Jews, grilled food consumption and rural residence for Arabs and Druze.

Geometric mean creatinine adjusted urinary concentrations of four hydroxyphenanthrenes (1OH\_phen, 2OH\_phen, 3OH\_phen, 4OH\_phen) were comparable to those found in the general US population aged 20 and older in 2007–2008 (CDC, 2013) whereas concentrations of 1OH\_pyrene were higher in our study population (140 ng/g compared to 80 ng/g in the US). On the other hand, urinary concentrations of 1OH\_phen, 2OH\_phen and 3OH\_phen were much lower in our study population compared to the general German

**Table 4**

PAHs urinary concentrations by ethnicity, Israeli adults, 2011 (N=243), results of linear regression model of log-transformed PAH values, adjusted for creatinine, cotinine and age.

Characteristics	1OH_pyrene		1OH_phen		2OH_phen <sup>a</sup>		3OH_phen		4OH_phen	
	R (95% CI) <sup>b</sup>	p-Value	R (95% CI)	p-Value	R (95% CI)	p-Value	R (95% CI)	p-Value	R (95% CI)	p-Value
Arabs and Druze vs. Jews—crude	0.99 (0.69; 1.40)	0.937	1.63 (1.16; 2.29)	0.005	1.50 (1.13; 1.98)	0.004	1.22 (0.87; 1.72)	0.243	1.65 (1.18; 2.30)	0.004
Arabs and Druze vs. Jews—adjusted for creatinine	1.02 (0.76; 1.38)	0.874	1.69 (1.26; 2.26)	<0.001	1.54 (1.21; 1.97)	0.002	1.27 (0.94; 1.70)	0.123	1.69 (1.23; 2.33)	0.002
Arabs and Druze vs. Jews—adjusted for cotinine	1.08 (0.77; 1.50)	0.653	1.81 (1.81; 2.46)	<0.001	1.64 (1.27; 2.11)	<0.001	1.38 (1.02; 1.86)	0.038	1.91 (1.44; 2.52)	<0.001
Arabs and Druze vs. Jews—adjusted for creatinine and cotinine	1.12 (0.86; 1.48)	0.394	1.89 (1.46; 2.44)	<0.001	1.69 (1.37; 2.09)	<0.001	1.43 (1.11; 1.84)	0.006	1.97 (1.52; 2.54)	<0.001
Arabs and Druze vs. Jews—adjusted for creatinine, cotinine and age (years)	1.13 (0.86; 1.49)	0.365	1.87 (1.44; 2.42)	<0.001	1.68 (1.36; 2.08)	<0.001	1.43 (1.11; 1.84)	0.006	1.95 (1.50; 2.52)	<0.001

<sup>a</sup> 2OH\_phen is the sum of 2- and 9-hydroxyphenanthrene.<sup>b</sup> Ratios (R) represent the multiplicative difference in PAH concentrations between Arab and Druze vs. Jewish study participants.**Table 5**

PAHs urinary concentrations (ng/L) by smoking characteristics and cotinine levels, Israeli adults, 2011 (N=243).

Variable	N	1OH_pyrene		1OH_phen		2OH_phen <sup>a</sup>		3OH_phen		4OH_phen	
		GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value
<b>Cotinine levels</b>											
<4 µg/L	140	130(106; 159)	<0.001	150(123;183)	<0.001	73(62; 87)	<0.001	94(77; 115)	<0.001	22(19;26)	<0.001
≥4 µg/L	103	251(206; 305)		263(216;320)		128(111; 148)		204(171;244)		42(34;53)	
<b>Smoking status<sup>b</sup></b>											
Non-smokers	144	135(110; 163)	0.001	154(127; 186)	0.002	76(64; 89)	<0.001	97(81; 118)	<0.001	23(19;27)	<0.001
Current smokers	90	233(187; 289)		247(197; 310)		119(100; 142)		189(154; 233)		39(30; 50)	
<b>Smoking mode (among current smokers)<sup>c</sup></b>											
Waterpipe only	13	205(122; 343)	0.629	162(110;239)	0.034	79(50;125)	0.055	130(75;225)	0.132	21(12;37)	0.037
Cigarettes	76	239(186; 308)		271(209;350)		129(106;157)		203(162;255)		44(34; 58)	
<b>Number of cigarettes/day (among cigarette smokers)</b>											
<10	24	170(108; 268)	Ref.	161(95; 273)	Ref.	86(61; 121)	Ref.	131(89; 195)	Ref.	21(14; 34)	Ref.
10–20	34	252(167; 380)	0.179	307(207; 455)	0.007	135(98; 187)	0.009	219(150; 321)	0.020	55(36; 83)	<0.001
>20	18	339(233; 495)	0.036	429(319; 575)	<0.001	204(165; 251)	<0.001	315(231; 429)	0.001	81(52; 126)	<0.001

<sup>a</sup> 2OH\_phen is the sum of 2- and 9-hydroxyphenanthrene.<sup>b</sup> Smoking status for 5 self-reported non-smokers with creatinine-adjusted cotinine levels above 150 µg/g was set to missing. Additional 4 participants were excluded due to unverified smoking status.<sup>c</sup> Data on smoking mode was missing for 1 participant.

**Table 6a**  
PAHs urinary concentrations (ng/L) by smoking characteristics and cotinine levels, Israeli adults, 2011, Jews (N = 183).

Variable	N	1OH_pyrene		1OH_phen		2OH_phen <sup>a</sup>		3OH_phen		4OH_phen	
		GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value
<b>Cotinine levels</b>											
<4 µg/L	101	121 (94; 155)	<0.001	122 (97; 154)	<0.001	62 (51; 76)	<0.001	83 (66; 105)	<0.001	18 (15; 22)	<0.001
≥4 µg/L	82	261 (210; 325)		249 (198; 314)		123 (105; 145)		201 (164; 245)		39 (30; 51)	
<b>Smoking status</b>											
Non-smokers	100	128 (100; 164)	0.001	126 (100; 158)	<0.001	65 (53; 79)	<0.001	88 (70; 111)	<0.001	19 (16; 23)	<0.001
Current smokers	75	231 (179; 297)		227 (175; 295)		110 (90; 134)		176 (139; 223)		35 (27; 46)	
<b>Smoking mode (among current smokers)<sup>b</sup></b>											
Waterpipe only	10	236 (127; 436)	0.958	160 (100; 257)	0.094	73 (42; 128)	0.106	123 (59; 256)	0.234	16 (9; 28)	0.021
Cigarettes	64	231 (174; 307)		245 (182; 329)		118 (95; 147)		187 (145; 242)		41 (30; 55)	
<b>Number of cigarettes/day (among cigarette smokers)</b>											
<10	22	171 (103; 282)	Ref.	155 (87; 276)	Ref.	84 (58; 122)	Ref.	131 (85; 201)	Ref.	21 (13; 34)	Ref.
10–20	30	236 (150; 373)	0.410	284 (184; 440)	0.025	123 (87; 175)	0.039	201 (132; 205)	0.075	49 (31; 7)	<0.001
>20	12	380 (222; 648)	0.060	390 (258; 583)	0.012	199 (153; 260)	0.001	304 (212; 436)	0.011	86 (55; 136)	<0.001
<b>Cotinine levels (among nonsmokers)</b>											
<LOQ (1 µg/L)	35	73 (48; 112)	0.001	78 (53; 115)	0.002	45 (32; 63)	0.006	53 (35; 80)	0.001	16 (12; 21)	0.135
≥LOQ (1 µg/L)	65	174 (131; 231)		163 (124; 214)		79 (62; 99)		115 (89; 150)		21 (16; 27)	

<sup>a</sup> 2OH\_phen is the sum of 2- and 9-hydroxyphenanthrene.

<sup>b</sup> Data on smoking mode was missing for 1 participant.

**Table 6b**  
PAHs urinary concentrations (ng/L) by smoking characteristics and cotinine levels, Israeli adults, 2011, Arabs and Druze (N = 56).

Variable	N	1OH_pyrene		1OH_phen		2OH_phen <sup>a</sup>		3OH_phen		4OH_phen	
		GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value	GM (95%CI)	p-Value
<b>Cotinine levels</b>											
<4 µg/L	36	152 (104; 223)	0.322	259 (180; 372)	0.486	118 (89; 157)	0.375	128 (87; 190)	0.108	38 (26; 55)	0.268
≥4 µg/L	20	207 (127; 337)		314 (218; 452)		144 (102; 205)		209 (137; 319)		54 (31; 93)	
<b>Smoking status</b>											
Non-smokers	41	144 (101; 206)	0.141	246 (177; 341)	0.190	110 (85; 143)	0.055	122 (86; 174)	0.022	36 (26; 51)	0.103
Current smokers	14	235 (154; 358)		365 (246; 541)		176 (125; 247)		261 (173; 395)		64 (33; 125)	
<b>Smoking mode (among current smokers)</b>											
Waterpipe only	3	128 (20; 847)	0.110	168 (28; 991)	0.098	104 (19; 569)	0.078	154 (49; 482)	0.156	52 (4; 784)	0.722
Cigarettes	11	276 (176; 432)		451 (318; 641)		203 (145; 284)		302 (186; 489)		69 (30; 156)	
<b>Number of cigarettes/day (among cigarette smokers)</b>											
<10	2	163 (20; 1355)	Ref.	242 (59; 1003)	Ref.	108 (9; 1338)	Ref.	136 (13; 1459)	Ref.	Unavailable due to small numbers	Ref.
10–20	4	403 (117; 1385)	0.032	540 (243; 1199)	0.014	275 (174; 435)	0.009	429 (228; 807)	0.020		0.099
>20	5	253 (121; 526)	0.177	501 (270; 929)	0.015	205 (111; 379)	0.039	313 (110; 891)	0.066		0.589
<b>Cotinine levels (among nonsmokers)</b>											
<LOQ (1 µg/L)	20	134 (78; 229)	0.692	256 (155; 424)	0.801	117 (77; 176)	0.653	114 (62; 209)	0.714	40 (24; 65)	0.591
≥LOQ (1 µg/L)	21	154 (91; 261)		236 (148; 376)		104 (73; 148)		130 (84; 202)		33 (19; 56)	

<sup>a</sup> 2OH\_phen is the sum of 2- and 9-hydroxyphenanthrene.

population in 1998 while 1OH<sub>2</sub>pyrene levels were higher in our population (140 ng/g compared to 110 ng/g) (Becker et al., 2003).

We found that urinary hydroxyphenanthrenes but not 1OH<sub>2</sub>pyrene concentrations were significantly higher in Jewish participants relative to Arab and Druze participants. After adjustment for age, creatinine and cotinine concentrations (mainly due to the adjustment for cotinine), associations accentuated, reaching a ratio of nearly two fold for 4OH<sub>2</sub>phen and 1OH<sub>2</sub>phen. The differences in urinary concentrations of PAHs by population sub-groups in Israel may be impacted by many factors, including exposure to ambient and indoor air pollutants, dietary preferences or lifestyle behaviors not captured by our study or differences in genetic factors impacting metabolism of these compounds, such as polymorphism in acetylators genes (Al-Daghri et al., 2013; Hansen et al., 2008). The exposure to PAHs by smoking among smokers might mask other predictors; hence analysis of risk factors among nonsmokers might elucidate mechanism of exposure.

There may be differences by population sub-groups in mode of cooking or consumption of specific food items not captured by our questionnaire. Such differences are hinted by our finding that grilled food consumption was a predictor of exposure among non-smoking Arabs and Druze but not among Jews. Practices such as burning biomass fuels such as wood on indoor open-pit stoves, which have been associated previously with high PAHs exposure (Chen et al., 2007; Hemat et al., 2012), may be more common among Arab and Druze populations in Israel. We were unable to directly explore this hypothesis, but our finding that rural residence was a risk factor for higher PAHs levels among nonsmoking Arabs and Druze, but not Jews, give indirect evidence to support this hypothesis. In contrast, higher cotinine levels, representing exposure to environmental tobacco smoking, were found in our study to be a risk factor for PAHs exposure among nonsmoking Jews but not Arabs and Druze, giving further support to differing modes of exposure by subpopulation groups. Evidence to higher exposure to PAHs among nonsmoking people exposed to environmental tobacco smoking, as manifested by higher cotinine levels, is accumulating (Aquilina et al., 2010; Polanska et al., 2014; Suwan-ampai et al., 2009). In our study such association was found for Jews but not for Arabs and Druze. This difference might be explained by other sources of exposure among nonsmoking Arabs and Druze, previously discussed such as diet, as manifested by relatively high levels of PAHs among nonsmoking Arabs and Druze with low cotinine levels. We were unable to explore other possible differences between Jews and Arabs exposure such as associations between PAHs levels in our study and exposure to ambient air pollution, distance from major roads, and distance for industrial PAHs sources.

Further study is needed to elucidate this finding and explore mechanisms of exposure to PAHs and possible impact on health among the Arab and Druze populations. Interestingly, previous analysis in our biomonitoring study revealed higher bisphenol A (BPA) exposure in the Jewish population (Berman et al., 2014), suggesting that ethnicity may be an important predictor of exposure to selected environmental contaminants in Israel. According to NHANES 1999–2002 data, household income level and to lesser extent race/ethnicity and education attainment, were significant determinants of urinary excretion of PAH metabolites (Suwan-ampai et al., 2009). There are limited data from previous studies on the associations between urinary PAHs levels and socio-demographic determinants, and no previous data among adults from the Middle East region.

Urinary excretions of all PAHs measured in the study were 1.5–2 times higher in participants reporting current smoking than in non-smokers. In addition, cotinine levels, representing total exposure to active and involuntary and involuntary tobacco smoke, were correlated with PAHs levels (Spearman correlation 0.35–0.45). However, correlation of cotinine with PAHs was weaker for Arabs and Druze,

implying that the relative contribution of tobacco smoking exposure to PAHs exposure is smaller in this group compared to Jews, i.e. another support for the hypothesis that there are additional sources of exposure, discussed before. The contribution of active cigarette smoking to PAHs exposure is well established. Multiple studies have consistently reported 1OH<sub>2</sub>pyrene concentration in urine to be 1.5–3 times higher among current smokers compared with nonsmokers, as well as for hydroxyphenanthrenes, although less frequently studied (Gündel et al., 1996; Suwan-ampai et al., 2009). Urinary PAHs levels in our study increased with number of cigarettes smoked per day. This association should not be taken for granted, as it was previously shown that the dose-response association between number of cigarettes and PAHs levels may differ between populations (Benowitz et al., 2011). Due to limited sample size we could not fully assess interaction between number of cigarettes smoked per day and PAHs levels by ethnicity, although the possibility of such difference by ethnicity was hinted by our data.

Levels of 1OH<sub>2</sub>pyrene, but not hydroxyphenanthrenes, among waterpipe only smokers were similar to cigarette smokers. This is, to the best of our knowledge, the first report of urinary levels of 1OH<sub>2</sub>pyrene among waterpipe smokers in an observational general population study. It was reported earlier that PAHs are emitted during waterpipe smoking (Shihadeh et al., 2012). This finding adds more evidence to the health risks of waterpipe smoking, to our previous findings of increased cotinine levels among exclusive waterpipe users (Levine et al., 2013) and increased utilization of health care services (Levine et al., 2012).

We found that grilled food consumption was a risk factor among nonsmoking Arabs and Druze, but we did not find associations with smoked food consumption or vegetarian status. In previous studies, differing by populations, PAHs measured and ages of participants, have found different dietary predictors of urinary PAHs levels including increased grilled fish and/or shellfish (Lee et al., 2009), consumption of grilled food and chocolate (Becker et al., 2007), consumption of grilled/boiled meat (Suwan-ampai et al., 2009). There are several possible explanations that our finding regarding dietary predictors of PAHs exposure was limited to non-smoking Arabs and Druze. First, while our dietary questionnaire included questions on overall consumption of smoked and grilled food we did not collect detailed information on consumption of specific foods, such as fish, cheese, or meat, chocolate, and well done foods. In addition, our limited sample size might have been underpowered to show weak associations or association with rare exposures. Among smokers, PAHs exposure from smoking might have masked other weaker or more rare dietary sources. Finally, we note that differences in dietary patterns across populations may explain differences in dietary predictors. For example, consumption of smoked and grilled foods might be low in Israel compared to other populations.

This study has several limitations. We employed a convenience non-random sampling technique to recruit individuals to the study. Therefore, it is unclear to what extent the study population is representative of the general adult population in Israel. The Druze population was over-represented in the study (17.9% compared to about 2% in the general population) whereas the non-Druze Arab population was under-represented. In addition, our estimates of dietary intake are based on self-reports of foods consumed over a 24 h period; although this is the method most commonly used for assessing short term food intakes, it must be recognized that the accuracy of data reported may be reduced by bias or error on the part of respondents or interviewers. These limitations, together with power limitation related to sample size, might explain lack of findings between dietary factors and PAHs levels.

On the other hand, this study had a number of strengths, including the fact that individuals were recruited from different ethnic

groups within Israel, and with wide geographical distribution. We directly measured cotinine levels which allowed us to better control for active and passive smoking exposure and avoid residual confounding. Finally, the laboratory methods employed gives a valid, sensitive and accurate quantitative biochemical assessment of personal PAHs exposure by a variety of metabolites.

Our finding regarding widespread public exposure to PAHs emphasizes the importance of reducing exposure in Israel, including interventions to reduce dietary, ambient, indoor and occupational exposures. In Israel, there are specific guidelines on limit values of PAHs in specific food items, and standards on ambient air levels of benzo(a)pyrene.

Within the framework of the national tobacco control plan, there have been important efforts to reduce active and second-hand smoking in Israel, although much more action is still needed. Human biomonitoring data, such as the data from the current study, can be instrumental in gaining support for further control measures and in directing services to susceptible populations, and can serve as baseline data to assess the impact of intervention.

In conclusion, we found that the general adult population in Israel is widely exposed to PAHs. The finding of higher exposure among Arabs and Druze highlights disparities in environmental exposures across subpopulations and suggests that further research and preventive measure are warranted to reduce PAHs exposure and associated health outcomes, especially in the Arab population in the Middle East.

## Acknowledgements

The study was supported by Research Grant Award No. RGA0902 from the Environment and Health Fund, Jerusalem, Israel.

## References

- Al-Daghri, N., Alokail, M., Abd-Alrahman, S., Draz, H., Yakout, S., Clerici, M., 2013. Polycyclic aromatic hydrocarbon exposure and pediatric asthma in children: a case-control study. *Environ. Health* 12, 1.
- Aquilina, N.J., Delgado-Saborit, J.M., Meddings, C., Baker, S., Harrison, R.M., Jacob III, P., Wilson, M., Yu, L., Duan, M., Benowitz, N.L., 2010. Environmental and biological monitoring of exposures to PAHs and ETS in the general population. *Environ. Int.* 36, 763–771.
- Becker, K., Conrad, A., Kirsch, N., Kolossa-Gehring, M., Schulz, C., Seiwert, M., Seifert, B., 2007. German Environmental Survey (GerES): human biomonitoring as a tool to identify exposure pathways. *Int. J. Hyg. Environ. Health* 210, 267–269.
- Becker, K., Schulz, C., Kaus, S., Seiwert, M., Seifert, B., 2003. German Environmental Survey 1998 (GerES III): environmental pollutants in the urine of the German population. *Int. J. Hyg. Environ. Health* 206, 15–24.
- Benowitz, N.L., Dains, K.M., Dempsey, D., Wilson, M., Jacob, P., 2011. Racial differences in the relationship between number of cigarettes smoked and nicotine and carcinogen exposure. *Nicotine Tob. Res.* 13, 772–783.
- Bergman, A., Heindel, J.J., Jobling, S., Kidd, K.A., Zoeller, R.T., Jobling, S.K., 2013. State of the Science of Endocrine Disrupting Chemicals 2012. An Assessment of the State of the Science of Endocrine Disruptors Prepared by a Group of Experts for the United Nations Environment Programme and World Health Organization.
- Berman, T., Goldsmith, R., Göen, T., Spungen, J., Novack, L., Levine, H., Amitai, Y., Shohat, T., Grotto, I., 2014. Demographic and dietary predictors of urinary bisphenol A concentrations in adults in Israel. *Int. J. Hyg. Environ. Health* 217, 638–644.
- Berman, T., Goldsmith, R., Göen, T., Spungen, J., Novack, L., Levine, H., Amitai, Y., Shohat, T., Grotto, I., 2013a. Urinary concentrations of environmental contaminants and phytoestrogens in adults in Israel. *Environ. Int.* 59, 478–484.
- Berman, T., Goldsmith, R., Göen, T., Spungen, J., Novack, L., Levine, H., Amitai, Y., Shohat, T., Grotto, I., 2013b. Urinary concentrations of organophosphate pesticide metabolites in adults in Israel: demographic and dietary predictors. *Environ. Int.* 60, 183–189.
- Boström, C.E., Gerde, P., Hanberg, A., Jernstrom, B., Johansson, C., Kyrklund, T., Ranng, A., Tornqvist, M., Victorin, K., Westerholm, R., 2002. Cancer risk assessment, indicators, and guidelines for polycyclic aromatic hydrocarbons in the ambient air. *Environ. Health Perspect.* 110 (Suppl 3), 451–488.
- CDC, U., 2013. Fourth National Report on Human Exposure to Environmental Chemicals—Updated Tables. CDC, U.
- Chen, B., Hu, Y., Jin, T., Zheng, L., Wang, Q., Shen, Y., Zhou, Y., 2007. Higher urinary 1-hydroxypyrene concentration is associated with cooking practice in a Chinese population. *Toxicol. Lett.* 171, 119–125.
- Eckert, E., Schmid, K., Schaller, B., Hiddemann-Koca, K., Drexler, H., Göen, T., 2011. Mercapturic acids as metabolites of alkylating substances in urine sample of German inhabitants. *Int. J. Hyg. Environ. Health* 214, 196–204.
- Feldt, T., Fobil, J.N., Wittsiepe, J., Wilhelm, M., Till, H., Zoufaly, A., Burchard, G., Göen, T., 2014. High levels of PAH-metabolites in urine of e-waste recycling workers from Agbogboshie, Ghana. *Sci. Total Environ.* 466, 369–376.
- Göen, T., Schaller, K.H., Drexler, H., 2012. External quality assessment of human biomonitoring in the range of environmental exposure levels. *Int. J. Hyg. Environ. Health* 215, 229–232.
- Gündel, J., Mannschreck, C., Büttner, K., Ewers, U., Angerer, J., 1996. Urinary levels of 1-hydroxypyrene, 1-, 2-, 3-, and 4-hydroxyphenanthrene in females living in an industrial area of Germany. *Arch. Environ. Contam. Toxicol.* 31, 585–590.
- Hansen, A.M., Mathiesen, L., Pedersen, M., Knudsen, L.E., 2008. Urinary 1-hydroxypyrene (1-HP) in environmental and occupational studies—a review. *Int. J. Hyg. Environ. Health* 211, 471–503.
- Heinrich, J., Holsscher, B., Seiwert, M., Carty, C., Merkel, G., Schulz, C., 2005. Nicotine and cotinine and adults' urine: the German environmental survey 1998. *J. Exposure Anal. Environ. Epidemiol.* 15, 74–80.
- Hemat, H., Wittsiepe, J., Wilhelm, M., Müller, J., Göen, T., 2012. High levels of 1-hydroxypyrene and hydroxyphenanthrenes in urine of children and adults from Afghanistan. *J. Exposure Sci. Environ. Epidemiol.* 22, 46–51.
- Hornung, R.W., Reed, L.D., 1990. Estimation of average concentration in the presence of nondetectable values. *Appl. Occup. Environ. Hyg.* 5, 46–51.
- Larsen, K., 1972. Creatinine assay by a reaction-kinetic principle. *Clin. Chim. Acta* 41, 209–217.
- Lee, K.-H., Vermeulen, R., Lenters, V., Cho, S.-H., Strickland, P.T., Kang, D., 2009. Determinants of urinary 1-hydroxypyrene glucuronide in South Korean children. *Int. Arch. Occup. Environ. Health* 82, 961–968.
- Levine, H., Berman, T., Goldsmith, R., Göen, T., Spungen, J., Novack, L., Amitai, Y., Shohat, T., Grotto, I., 2013. Exposure to tobacco smoke based on urinary cotinine levels among Israeli smoking and nonsmoking adults: a cross-sectional analysis of the first Israeli human biomonitoring study. *BMC Public Health* 13, 1241.
- Levine, H., Huerta-Hartal, M., Bar-Ze'ev, Y., Balicer, R., Auster, O., Ankol, O., Chodick, G., 2012. Smoking and other correlates of health care services utilization among mandatory military recruits in Israel. *Nicotine Tob. Res.* 14, 742–750.
- Li, Z., Sandau, C., Romanoff, L., Caudill, S., Sjodin, A., Needham, L., Patterson, D., 2008. Concentration and profile of 22 urinary polycyclic aromatic hydrocarbon metabolites in the US population. *Environ. Res.* 107, 320–331.
- Müller, M., Heinrich-Ramm, R., Hoppe, H.W., 2003. Working group Analyses in Biological Materials, Cotinine – Determination in urine to ascertain passive exposure to smoking. In: Angerer, J., Schaller, K.H., Greim, H. (Eds.), *Analyses of Hazardous Substances in Biological Materials*, 8. Wiley-VCH, Weinheim, pp. 53–65. <http://dx.doi.org/10.1002/3527600418.bi48656e0008>.
- Perera, F.P., Tang, D., Wang, S., Vishnevetsky, J., Zhang, B., Diaz, D., Camann, D., Rauh, V., 2012. Prenatal polycyclic aromatic hydrocarbon (PAH) exposure and child behavior at age 6–7 years. *Environ. Health Perspect.* 120, 921.
- Phillips, D.H., 1999. Polycyclic aromatic hydrocarbons in the diet. *Mutat. Res.* 443, 139–147.
- Polanska, K., Hanke, W., Dettbarn, G., Sobala, W., Gromadzinska, J., Magnus, P., Seidel, A., 2014. The determination of polycyclic aromatic hydrocarbons in the urine of non-smoking Polish pregnant women. *Sci. Total Environ.* 487, 102–109.
- Schulz, C., Conrad, A., Becker, K., Kolossa-Gehring, M., Seiwert, M., Seifert, B., 2007. Twenty years of the German Environmental Survey (GerES): human biomonitoring—temporal and spatial (West Germany/East Germany) differences in population exposure. *Int. J. Hyg. Environ. Health* 210, 271–297.
- Shihadeh, A., Salman, R., Jaroudi, E., Saliba, N., Sepetdjian, E., Blank, M.D., Cobb, C.O., Eissenberg, T., 2012. Does switching to a tobacco-free waterpipe product reduce toxicant intake? A crossover study comparing CO, NO, PAH, volatile aldehydes, tar and nicotine yields. *Food Chem. Toxicol.* 50, 1494–1498.
- Straif, K., Baan, R., Grosse, Y., Secretan, B., El Ghissassi, F., Coglian, V., 2005. Carcinogenicity of polycyclic aromatic hydrocarbons. *Lancet Oncol.* 6, 931–932.
- Suwan-ampai, P., Navas-Acien, A., Strickland, P.T., Agnew, J., 2009. Involuntary tobacco smoke exposure and urinary levels of polycyclic aromatic hydrocarbons in the United States, 1999 to 2002. *Cancer Epidemiol. Biomarkers Prevent.* 18, 884–893.